

## IMPACT OF PRESCRIBED FIRE SMOKE ON DOWNWIND COMMUNITIES

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### Abstract

Prescribed fires are conducted each year across Australia for fuel management and/or ecological reasons. In some regions, during the period of intensive prescribed burning, large quantities of hazardous air contaminants may be emitted and may exceed urban air quality guidelines. The impact on community health will depend on what hazardous pollutants the population is exposed to, the levels of exposures and the potential for such exposures to cause adverse health effects among the community.

The paper will discuss findings from a review of Australian and international literature regarding implications of exposure to bushfire smoke 'air toxics' on community health. Most studies have been on large fires, whether accidental or during forest clearing activities, with little to no research on community health and air toxics exposures downwind of prescribed fires. The review has shown that the primary pollutant consistently exceeding air quality guidelines downwind of large bushfires was particulate matter, but it also highlighted that there is clearly a need to further investigate the effects of bushfire on public health. Some of the issues to be addressed include

- Monitoring the seasonal exposure of communities to hazardous pollutants released during prescribed burning activities and determining major factors that influence exposure levels
- Assessing the health implications of such exposures, especially in relation to existing (urban) air quality measures or by developing a bushfire air quality index as a tool for better risk management.

### Introduction

Bushfires, both accidental and planned, occur each year in many parts of the world. These bushfires can range from small-scale to large-scale, can burn for hours or weeks, and can release substantial quantities of hazardous air contaminants ('air toxics') into the air environments of communities. In Australia, the practice is to conduct fuel reduction burns during spring and autumn to remove fine, flammable fuels and thereby reduce the severity of unplanned fires. During the period of intensive burning activities, bushfire smoke can travel substantial distances to surrounding rural towns, resulting in regular annual exposures to rural communities. The impact on community health will depend on *what* hazardous pollutants the communities are exposed to, the *levels of exposure* to these pollutants, and whether *adverse health effects* are likely to occur in the *populations exposed*.

While much data are available on urban exposures to hazardous pollutants, there is currently a lack of measurements of these pollutants in rural areas. Rural communities undergo seasonal exposures to toxic air pollutants as a result of prescribed burns, but little is known of the levels of exposure that occur and potential adverse health impacts. This review will present findings on community exposures downwind from major bushfire events, and will consider exposure assessment, health impacts and exposure criteria.

### Community Exposure Standards and Guidelines for Urban Air Pollutants

It is well known that urban communities are exposed to a range of air pollutants associated with urban activities, especially automobile exhausts. Regulations to control these exposures exist in Australia and most developed countries for what are termed 'criteria pollutants'. In Australia, these are applied as National Environmental Protection Measures (NEPMs) for ambient air quality, aimed at protecting the *health and well-being* of urban populations, as summarised in Table 1 (National Environmental Protection Council 2003). In Australia, the current ambient air NEPMs (24-hour) are  $50 \mu\text{g}/\text{m}^3$  for  $\text{PM}_{10}$  (particulate matter with aerodynamic diameter less than  $10 \mu\text{m}$ ) and  $25 \mu\text{g}/\text{m}^3$  for  $\text{PM}_{2.5}$  (particulate matter with aerodynamic diameter less than  $2.5 \mu\text{m}$ ). It will be shown later in this review that such levels are often greatly exceeded for communities living downwind of large bushfires, though *it is unknown whether urban particles and bushfire smoke have equivalent health impacts at these exposures*.

### Bushfire Emissions and Effects on Community Health

During bushfires, a wide range of air contaminants are released, which may cause adverse health effects if communities are exposed to elevated concentrations for extended periods. Air toxics in bushfire smoke are present in both the particle and gaseous phases, and include respirable particles, carbon monoxide (CO), carbon dioxide, nitrogen- and sulfur-based compounds, aldehydes, volatile and semi-volatile organic compounds (VOCs and SVOCs), polycyclic

aromatic hydrocarbons (PAHs), dioxins, organic acids, free radicals and ozone (O<sub>3</sub>) (Brauer 1999; Fujiwara *et al.* 1999; Malilay 1999; Ward 1999).

**Table 1: National Environmental Protection Measures (NEPMs) for ambient air quality**

Air contaminant	Average period	Maximum concentration	Maximum allowable exceedances
Carbon monoxide	8 hours	9.0 ppm	1 day per year
Nitrogen dioxide	1 hour	0.12 ppm	1 day per year
	1 year	0.03 ppm	None
Photochemical oxidant (as ozone)	1 hour	0.10 ppm	1 day per year
	4 hours	0.08 ppm	1 day per year
Sulfur dioxide	1 hour	0.2 ppm	1 day per year
	1 day	0.08 ppm	1 day per year
	1 year	0.02 ppm	None
Lead	1 year	0.5 µg/m <sup>3</sup>	None
Particles as PM <sub>10</sub>	1 day	50 µg/m <sup>3</sup>	5 days per year
Particles as PM <sub>2.5</sub>	1 day	25 µg/m <sup>3</sup>	Gather sufficient data for review in 2005
	1 year	8 µg/m <sup>3</sup>	
Benzene	1 year	0.003 ppm	Gather sufficient data to develop standard in 2009
Toluene	1 day	1.0 ppm	Gather sufficient data to develop standard in 2009
	1 year	0.10 ppm	
Xylenes	1 day	0.25 ppm	Gather sufficient data to develop standard in 2009
	1 year	0.20 ppm	
Formaldehyde	1 day	0.04 ppm	Gather sufficient data to develop standard in 2009
Benzo(a)pyrene	1 year	0.3 ng/m <sup>3</sup>	Gather sufficient data to develop standard in 2009

Many studies have investigated the relationship between bushfire smoke exposure to downwind communities and a range of health outcomes such as mortality records, hospital admissions, emergency department visits for asthma or pulmonary disease, asthma exacerbation, change in lung function, increase of cardiopulmonary symptoms (coughing, wheezing, shortness of breath, angina), upper respiratory illness, as well as mucous membrane irritation. These studies relied primarily on hospital-based records (e.g. number of emergency department visits and hospital admissions during bushfire episodes), but very few community-based studies (e.g. epidemiological studies of communities) were conducted other than community surveys. Furthermore the majority of the studies focused on particulate matter, which was the primary pollutant that consistently exceeded air quality guidelines. This is a significant finding, since fine particles in urban air (not from bushfires) have been found to impact the health of urban communities. A summary of the reviewed studies is presented in Table 2.

The studies show several consistent findings:

- Bushfires generated large amounts of particulate matter, especially respirable particles, which greatly exceeded air quality standards. Maximum daily PM<sub>10</sub> were recorded at approximately 200–1000 µg/m<sup>3</sup>; any daily PM<sub>10</sub> above 100–200 µg/m<sup>3</sup> was considered a high pollution episode.
- Other criteria pollutants (O<sub>3</sub>, NO<sub>x</sub>, SO<sub>2</sub> and CO) may have increased, but generally stayed within the health-based environmental standards. Elevated levels of CO and PAHs were observed in Indonesia during the 1997 fires (Kunii 1999; Aditama 2000; Kunii *et al.* 2002).
- Adverse health impacts were observed for all smoke exposures, except in two studies in suburban Sydney (Cooper *et al.* 1994; Smith *et al.* 1996). The health impacts were most severe for the Southeast Asian fires that involved the highest PM levels and the longest exposure periods.
- Adverse health impacts were found for: mortality rates in some studies; hospital and emergency department admissions for asthma or pulmonary disease; asthma exacerbation; reduced lung function; increase of cardiopulmonary symptoms (coughing, wheezing, shortness of breath, heart problems); upper respiratory illness; and mucous membrane irritation. Usually the observed health impacts were linked to PM levels on the basis of their high elevation compared to health-based environmental exposure standards.
- Within the significant limitations of different fire types, fire durations and PM measurements presented above, significant health effects were observed at levels that greatly exceed the current Australian one-day NEPMs for

PM<sub>10</sub> and PM<sub>2.5</sub> of 50 and 25 µg/m<sup>3</sup>, respectively. No data are available to confirm the adequacy of these standards for the protection of communities exposed to bushfire smoke.

- Indoor levels of fine particles varied little from outdoor levels downwind of the bushfires (Kunii *et al.* 2002; Radojevic 2003; Sapkota *et al.* 2005), indicating that deposition losses of smoke particles to building surfaces were small and resulted in little improvement of protection from bushfire smoke.

**Table 2. Studies of PM exposures and health impacts on communities downwind of bushfires**

	Air Toxics (Units PM [µg/m <sup>3</sup> ])	Increased health effects <sup>A</sup>	Reference
United States	Max. PM <sub>10</sub> >150	A, COPD	Duclos <i>et al.</i> 1990
	Not measured	Bronchospastic and irritative reactions	Shusterman <i>et al.</i> 1993
	Not measured	A, B, chest pain	US Centers for Disease Control and Prevention 1999
	PM <sub>10</sub> max. >1000, 16 day >150, 2 day >500	A, COPD and RD	Mott and Meyer 2000; Mott <i>et al.</i> 2002
	PM <sub>2.5</sub> max., ave. >65, no increases for PAHs and benzene; CO below std	RD, H	Ward and Smith 2001; Bible 2002
	PM <sub>2.5</sub> 63; CO 1 ppm	Worsening symptoms in 21 COPD	Sutherland <i>et al.</i> 2005
	PM <sub>2.5</sub> max. 90-200	Not measured	Sapkota <i>et al.</i> 2005
	PM <sub>10</sub> max. 215; increases for CO and NO, but not for NO <sub>2</sub> and O <sub>3</sub>	Not measured	Phuleria <i>et al.</i> 2005
Australia	Nephelometer readings	A	Churches 1991
	PM <sub>10</sub> max. 250, median 18; no difference for O <sub>3</sub> and NO <sub>2</sub>	No increase in A or RD	Cooper <i>et al.</i> 1994; Smith <i>et al.</i> 1996
	PM <sub>10</sub> max. 210	Peak exp. flow rates in children (wheeze), no change	Jalaludin <i>et al.</i> 2000
	Max. PM <sub>2.5</sub> 70	A esp when PM <sub>2.5</sub> > 40	Johnston <i>et al.</i> 2002
Southeast Asia	PM <sub>10</sub> max. 1800; CO very unhealthful; SO <sub>2</sub> , NO <sub>2</sub> and O <sub>3</sub> good to moderate; increase in PAHs	A, RI, RD, pneumonia and worsening of respiratory symptoms, mortality (resp. failure)	Kunii 1999; Aditama 2000; Frankenberg <i>et al.</i> 2002; Kunii <i>et al.</i> 2002
	PM <sub>10</sub> max. 930; High poll. day when PM <sub>10</sub> >210	A, RI, RD, COPD (>65 yo); elderly and people with A had larger effect; mortality (>65 yo); decreased lung function among school children	Brauer and Hisham-Hashim 1998; Hisham-Hashim <i>et al.</i> 1998; WHO 1998; Awang <i>et al.</i> 2000; Sastry 2000; Mott <i>et al.</i> 2005
	PM <sub>10</sub> range 44-60	A (<12 yo) correlated with PM <sub>10</sub> ; higher frequency of asthmatic attacks	Chew <i>et al.</i> 1995; Chia <i>et al.</i> 1995
	PM <sub>10</sub> doubled to 60-100; slight increase for CO, NO <sub>2</sub> and O <sub>3</sub> ; no change for SO <sub>2</sub>	A, R; mortality, no change	Emmanuel 2000
	Monthly PM <sub>10</sub> increased from 48 to 69; max. daily 218; no change for NO <sub>2</sub> and SO <sub>2</sub>	RD, B, COPD, A	Phonboon <i>et al.</i> 1999
	PM <sub>10</sub> max. 1000; avg 110; no change for NO, NO <sub>2</sub> and O <sub>3</sub> ; SO <sub>2</sub> , 4x increase; CO, 10x increase	Higher freq. of adverse health impacts for 1-5 yo, >60 yo and outdoor workers	Muraleedharan <i>et al.</i> 2000; Odihi 2001; Radojevic 2003

<sup>A</sup> Increased medical incidences of asthma (A), respiratory disease (RD), respiratory infection (RI), heart (H), chronic obstructive pulmonary disease (COPD), bronchitis (B) and rhinitis (R)

While many studies found that the levels of fine particulate matter were consistently and highly elevated downwind of bushfires, levels of other pollutants often varied little from baseline levels in urban areas, or showed increases that were much lower than health-based environmental standards. This difference is considered to reflect the different yields of pollutants, their dispersion in buoyant plumes and by meteorological effects over large distances, and the different environmental standards for pollutants. Smoke particles have a high yield in combustion and a (numerically) low environmental standard has been set to protect the general population from adverse health effects of urban particles. While bushfire smoke and urban particles have similar sizes, there may be significant differences in the constituents attached to the particles and their interaction with other pollutant species present in the smoke, and therefore they may

affect community health differently. The studies reviewed above certainly identify similar morbidity impacts for urban particulates and bushfire smoke, and there is limited evidence for similar mortality effects, but *the exposure and health impacts are poorly quantified for bushfire smoke*. An environmental standard specific to bushfire smoke particles may be more appropriate.

### **Community Response to Bushfire Air Toxics**

Communities can react in different ways to bushfire smoke to reduce their levels of exposure to air toxics. Government bodies may advise them to wear respiratory protection, to reduce physical exertion and to remain indoors. In response to fire danger, they may be advised (or required) to evacuate an area, though Australian fire agencies generally recommend home occupants remain under a 'stay or go' plan, where they either leave early or stay if well prepared to protect themselves and their homes. Assuming that home occupants stay in the bushfire smoke, their options to reduce air toxics exposures are:

- *Wear a respirator with a suitable protection factor* for the levels of air toxics, a strategy which was adopted in Malaysia (Hisham-Hashim *et al.* 1998). Assuming from this review that fine particle exposures are the primary concern, a particle respirator with a 10-fold protection factor would reduce exposure to one-tenth of the smoke levels. This level of protection has been selected since it is the maximum that can be provided by half-face respirators that are moderately comfortable to wear (Standards Australia and Standards New Zealand 1994)
- *Stay indoors with external doors and windows shut*. The basis for this recommendation is that short-term peak smoke levels will be avoided, and that particulate matter is partially deposited to surfaces as it enters a building with ventilation air (a loss we estimate to be ~20%). Note that recent research in Canadian communities downwind of bushfires determined a median indoor-to-outdoor pollutant concentration ratio of 0.91, indicating that staying indoors did not protect people from smoke exposure (Sapkota *et al.* 2005).

During the Southeast Asian fires, a Pollutant Standards Index (PSI) from the USA was used to characterize smoke haze events, which takes into account the five criteria pollutants (PM<sub>10</sub>, CO, NO<sub>2</sub>, SO<sub>2</sub> and O<sub>3</sub>). PSI levels above 100 indicate that the ambient concentrations exceed at least one of the air quality standards and triggers preventative action by government, which could include health advice warnings. However the PSI ignores many of the air toxics of Table 1 that may be present in bushfire smoke, limiting the applicability of this index for bushfire haze episodes. A broader index may be appropriate for bushfire smoke.

### **Conclusions**

More detail of each of these studies is presented elsewhere (Reisen and Brown 2006) but significant observations on further research needs are presented here. The studies in this review covered major bushfire events, predominantly overseas. They have shown that bushfire smoke consistently caused respirable particle levels in downwind communities to exceed ambient air quality standards, and that this was generally considered to be the primary factor in the adverse health effects observed in communities. However, assigning the health effects to particles assumes bushfire particles and urban air particles have similar health impacts and standards. Exposures to CO, NO<sub>2</sub>, SO<sub>2</sub> or O<sub>3</sub> were increased in bushfire smoke, but not often above standards. Overall, little attention was given to the impact of PAHs, VOCs and aldehydes on community health and to the potential for interactive effects of pollutants.

Some studies observed increased mortality, but most observed increased morbidity, primarily among the susceptible population, which includes children, the elderly and people with asthma and pre-existing respiratory and/or cardiovascular disease. In general, healthy adults quickly recovered from short-term bushfire smoke exposures. An association between bushfire episodes and health impacts, measured by hospital visits for asthma and respiratory illnesses, were observed in South-East Asia and the USA. However, out of three Australian studies, only Johnston *et al.* (2002) reported an adverse health outcome as a result of bushfires around Darwin based on asthma attendances to major hospitals. No explanation is available for this significant difference in findings. There is clearly a need in Australia to further investigate the effects of bushfires on community health.

Even though the studies have all been related to accidental and forest clearing bushfires, it is possible that similar effects of bushfire smoke impact on community health will be observed during a season of intensive prescribed burn activities. In order to determine this potential, monitoring of the air toxics during prescribed burn seasons should be carried out for downwind communities so that the major factors that affect exposure levels are determined. Even in the reports of large fires reviewed in this paper, the determinants of exposure levels have not been sought and these studies have provided little to no information on the fuels combusted in the fires. This could be a key factor in the amount and

nature of the pollutants formed, but is largely ignored in discussion. Other important details of the fires were also often lacking, e.g. fire conditions and proximity to communities, as well as meteorological conditions (wind speed, inversion).

Indoor measurements of bushfire air toxics have also been very limited, but generally indicate that staying indoors does not protect people from smoke exposure (other than short-term peaks). There is a need for research on how to protect indoor occupants from bushfire smoke.

## References

- Aditama, T. Y. (2000). Impact of haze from forest fire to respiratory health: Indonesian experience. *Respirology*, **5**, 169-174.
- Awang, M. B., Jaafar, A. B., Abdullah, A. M., Ismail, M. B., Hassan, M. N., Abdullah, R., Johan, S. and Noor, H. (2000). Air quality in Malaysia: impacts, management issues and future challenges. *Respirology*, **5**, 183-196.
- Bible, R. (2002). Breathless. *Wildfire Magazine*, **July/August**.
- Brauer, M. (1999). Health impacts of biomass air pollution. In "Health Guidelines for Vegetation Fire Events: Background Papers". p.186-255. World Health Organization, Geneva.
- Brauer, M. and Hisham-Hashim, J. (1998). FIRES in Indonesia: Crisis and reaction. *Environmental Science & Technology*, **32**(17), 404a-407a.
- Chew, F. T., Ooi, B. C., Hui, J. K. S., Saharom, R., Goh, D. Y. T. and Lee, B. W. (1995). Singapore Haze and Acute Asthma in Children. *Lancet*, **346**(8987), 1427-1427.
- Chia, H. P., Chia, K. S., Ooi, P. L., Ng, T. P., Goh, K. T. and Lee, H. P. (1995). Effects of the recent haze in Singapore on the frequency of attacks among group of known asthmatics. In "Health and the Built Environment". p. 87-93. Institute of Environmental Epidemiology, Singapore.
- Churches, T. (1991). Asthma and air pollution in Sydney. *NSW Public Health Bulletin*, **2**(8), 72-73.
- Cooper, C. W., Mira, M., Danforth, M., Abraham, K., Fasher, B. and Bolton, P. (1994). Acute Exacerbations of Asthma and Bushfires. *Lancet*, **343**(8911), 1509-1509.
- Duclos, P., Sanderson, L. M. and Lipsett, M. (1990). The 1987 Forest Fire Disaster in California - Assessment of Emergency Room Visits. *Archives of Environmental Health*, **45**(1), 53-58.
- Emmanuel, S. C. (2000). Impact to lung health of haze from forest fires: the Singapore experience. *Respirology*, **5**, 175-182.
- Frankenberg, E., McKee, D. and Thomas, D. (2002). Health consequences of forest fires in Indonesia. *IUSSP Regional Population Conference*, Bangkok.
- Fujiwara, M., Kita, K., Kawakami, S., Ogawa, T., Komala, N., Saraspriya, S. and Suripto, A. (1999). Tropospheric ozone enhancements during the Indonesian forest fire events in 1994 and in 1997 as revealed by ground-based observations. *Geophysical Research Letters*, **26**(16), 2417-2420.
- Hisham-Hashim, Hashim, J. Z., Jalaludin, J., Lubis, S. and Hashim, R. (1998). Respiratory function of elementary school children exposed to the 1997 Kuala Lumpur haze. *Epidemiology*, **9**(4), S103.
- Jalaludin, B., Smith, M., O'Toole, B. and Leeder, S. (2000). Acute effects of bushfires on peak expiratory flow rates in children with wheeze: a time series analysis. *Australian and New Zealand Journal of Public Health*, **24**(2), 174-177.
- Johnston, F. H., Kavanagh, A. M., Bowman, D. M. J. S. and Scott, R. K. (2002). Exposure to bushfire smoke and asthma: an ecological study. *Medical Journal of Australia*, **176**(11), 535.
- Kunii, O. (1999). Basic facts - determining downwind exposures and their associated health effects, assessment of health effects in practice: a case study in the 1997 forest fires in Indonesia. In "Health Guidelines for Vegetation Fire Events: Background Papers". p. 295-312. World Health Organization, Geneva.
- Kunii, O., Kanagawa, S., Yajima, I., Hisamatsu, Y., Yamamura, S., Amagai, T. and Ismail, I. T. S. (2002). The 1997 haze disaster in Indonesia: Its air quality and health effects. *Archives of Environmental Health*, **57**(1), 16-22.
- Malilay, J. (1999). A review of factors affecting the human health impacts of air pollutants from forest fires. In "Health Guidelines for Vegetation Fire Events: Background Papers". p. 255-270. World Health Organization, Geneva.
- Mott, J. and Meyer, P. (2000). Health effects associated with forest fires among residents of the Hoopa Valley National Indian Reservation. Air Pollution and Respiratory Health Branch, CDC, Atlanta, GA.
- Mott, J. A., Mannino, D. M., Alverson, C. J., Kiyu, A., Hashim, J., Lee, T., Falter, K. and Redd, S. C. (2005). Cardiorespiratory hospitalizations associated with smoke exposure during the 1997 Southeast Asian forest fires. *International Journal of Hygiene and Environmental Health*, **208**(1-2), 75-85.
- Mott, J. A., Meyer, P., Mannino, D., Redd, S. C., Smith, E. M., Gotway-Crawford, C. and Chase, E. (2002). Wildland forest fire smoke: health effects and intervention evaluation, Hoopa, California, 1999. *Western Journal of Medicine*, **176**(3), 157-162.
- Muraleedharan, T. R., Radojevic, M., Waugh, A. and Caruana, A. (2000). Chemical characterisation of the haze in Brunei Darussalam during the 1998 episode. *Atmospheric Environment*, **34**(17), 2725-2731.
- National Environmental Protection Council (2003). National Environment Protection (Ambient Air Quality) Measure (Amendment). Canberra, EPHC.
- Odihi, J. O. (2001). Haze and health in Brunei Darussalam: The case of the 1997-98 episodes. *Singapore Journal of Tropical Geography*, **22**(1), 38-51.
- Phonboon, K., Paisarn-uchapong, O., Kanatharan, P. and Agsrn, S. (1999). Smoke episodes emissions characterization and assessment of health risks related to downwind air quality - case study, Thailand. In "Health Guidelines for Vegetation Fire Events: Background Papers". p. 334-380. World Health Organization, Geneva.
- Phuleria, H. C., Fine, P. M., Zhu, Y. F. and Sioutas, C. (2005). Air quality impacts of the October 2003 Southern California wildfires. *Journal of Geophysical Research-Atmospheres*, **110**(D7), -.
- Radojevic, M. (2003). Haze research in Brunei Darussalam during the 1998 episode. *Pure and Applied Geophysics*, **160**(1-2), 251-264.

- Reisen, F. and Brown, S. K. (2006). Implications for community health from exposure to bushfire air toxics. *Environmental Chemistry*, submitted.
- Sapkota, A., Symons, J. M., Kleissl, J., Wang, L., Parlange, M. B., Ondov, J., Breysse, P. N., Diette, G. B., Eggleston, P. A. and Buckley, T. J. (2005). Impact of the 2002 Canadian forest fires on particulate matter air quality in Baltimore City. *Environmental Science & Technology*, **39**(1), 24-32.
- Sastry, N. (2000). Forest Fires, Air Pollution and Mortality in Southeast Asia. RAND Labor and Population Program, Santa Monica, CA.
- Shusterman, D., Kaplan, J. Z. and Canabarro, C. (1993). Immediate Health-Effects of an Urban Wildfire. *Western Journal of Medicine*, **158**(2), 133-138.
- Smith, M. A., Jalaludin, B., Byles, J. E., Lim, L. and Leeder, S. R. (1996). Asthma presentations to emergency departments in western Sydney during the January 1994 bushfires. *International Journal of Epidemiology*, **25**(6), 1227-1236.
- Standards Australia and Standards New Zealand (1994). Selection, Use and Maintenance of Respiratory Protective Devices, AS/NZS 1715:1994 (Standards Australia: Sydney).
- Sutherland, E. R., Make, B. J., Vedal, S., Zhang, L. N., Dutton, S. J., Murphy, J. R. and Silkoff, P. E. (2005). Wildfire smoke and respiratory symptoms in patients with chronic obstructive pulmonary disease. *Journal of Allergy and Clinical Immunology*, **115**(2), 420-422.
- US Centers for Disease Control and Prevention (1999). Surveillance of morbidity during wildfires - central Florida, 1998. *Morb. Mortal, Weekly Rep.*, **48**(4), 78-79.
- Ward, D. E. (1999). Smoke from wildland fires. In "Health Guidelines for Vegetation Fire Events: Background Papers". p. 70-85. World Health Organization, Geneva.
- Ward, T. J. and Smith, G. C. (2001). Air sampling study of the 2000 Montana wildfire season. *Air and Waste Management Conference*, Orlando, FL.
- WHO (1998). Biregional Workshop on Health Impacts of Haze-Related Air Pollution. World Health Organization, Regional Offices for Southeast Asia and the Western Pacific, Kuala Lumpur, Malaysia.